

1 **Main Manuscript for**

2 **Potential Impacts of Fukushima Nuclear Wastewater Discharge on Nutrient**
3 **Supply and Greenhouse Gas Emissions of Food Systems.**

4

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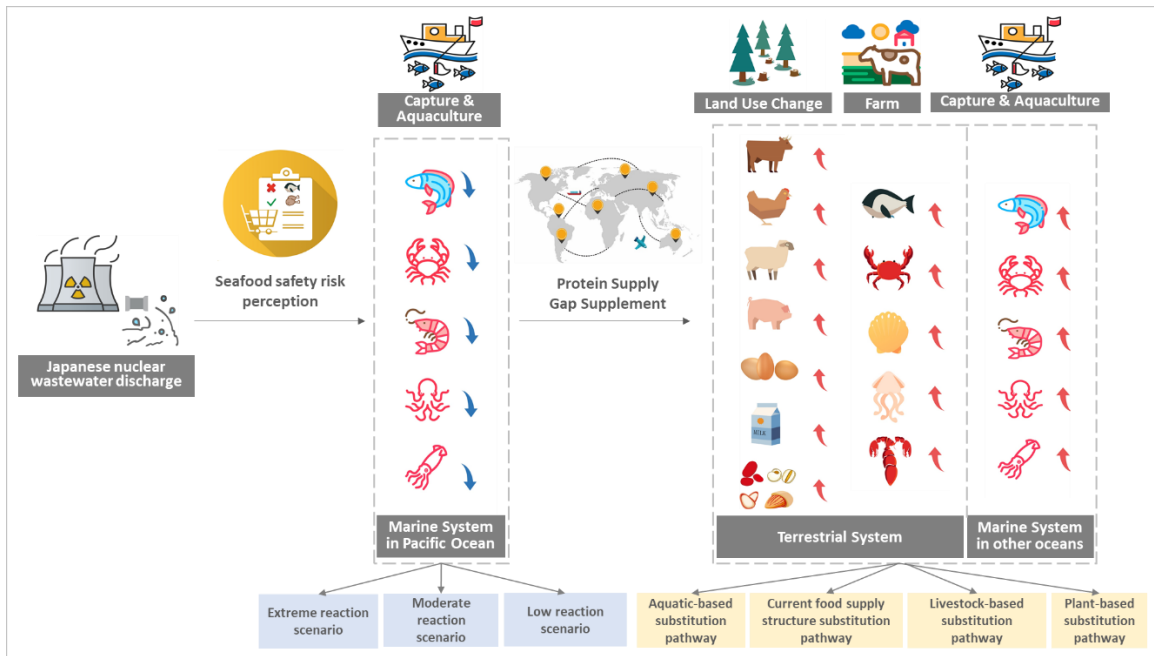
31 **Abstract**

32 Seafood plays a critical role in promoting a nutritious and climate-friendly food
33 system but faces challenges from marine contamination that pose food safety risks. To
34 address these challenges, we developed a Marine Pollution-induced Food System Risk
35 Assessment and Response (MP-FSRAR) model to assess the impact of consumer
36 perceptions of marine pollution risk events on nutrient supply and greenhouse gas (GHG)
37 emissions. Applying this model to the proposed Fukushima Nuclear Wastewater Discharge
38 (FNWD) into the Pacific Ocean, we identified the potential variation in nutrient supply and
39 GHG emissions resulting from changes in Chinese consumers' willingness to consume
40 seafood under 12 scenarios. Our analysis revealed nutrient supply gaps in daily protein,
41 minerals, vitamins, and fatty acids, ranging from 1.5-2.7 g/person, 2.5-4.5 mg/person, 5.6-
42 10.3 mcg/person, and 0.08-0.14 g/person, respectively. Closing the protein gap in China
43 would require an additional 3.2-15.8 billion kg yr⁻¹ protein-rich foods, including aquatic,
44 livestock, dairy, and pulses. Additional food supply would increase extra GHG emissions
45 by 1-79 billion kg CO₂e yr⁻¹, resulting in potential GHG emissions being transmitted to
46 other countries through the trade chain. Furthermore, we found that the livestock-based
47 food substitution pathway had 40 times higher GHG emissions than the plant-based one.
48 Our MP-FSRAR model provides global stakeholders with new perspectives on the
49 potential consequences of consumer perceptions of marine pollution risks and offers
50 valuable insights for developing countermeasures to build a more resilient and low-carbon
51 food system.

52

53

54 **Graphical Abstract**



55

56 **Keywords:** Marine pollution, consumer perceptions, food security, nutrition supply,
57 GHG emissions.

58 **1. Introduction**

59 Global food demand is expected to rise to an unprecedented extent for the following
60 decades, propelled by a projected one-third increase in total population and higher average
61 individual incomes by 2050 (Godfray et al., 2010; Mueller et al., 2012; Tilman, Balzer,
62 Hill, & Befort, 2011). Unfortunately, the current reality is that over one in four people suffer
63 from hunger or lack regular access to nutritious and adequate food (Organization, 2020).
64 To make matters worse, the prolonged COVID-19 pandemic, climate change, and regional
65 conflicts pose additional stronger headwinds to global food security (Fujimori et al., 2019;
66 Laborde, Martin, Swinnen, & Vos, 2020; Nature, 2022). Moreover, the current food system
67 significantly contributes to global anthropogenic GHG emissions, with approximately one-
68 third coming from meat consumption (Crippa et al., 2021; Foley et al., 2011; J Poore & T

69 Nemecek, 2018). Such food system emissions could preclude the attainment of 1.5°C-2°C
70 global climate change targets (Clark et al., 2020). The world is thus facing the formidable
71 double challenge of ensuring the continued delivery of sufficient food while curbing overall
72 GHG emissions.

73 Momentum is gathering to seek food from the ocean for a more nutritious and climate-
74 friendly food system (Hicks et al., 2019; Walter Willett et al., 2019), leading to a significant
75 increase in global capture fisheries and aquaculture production in recent decades (Fig. S1).
76 Seafood consumption provides more than 15% of the average animal protein intake for 2.9
77 billion people (Smith Martin et al., 2010; Walter Willett et al., 2019), as well as being a
78 vital source of essential micronutrients such as amino acids, omega-3 fatty acids, vitamins,
79 and minerals, which are essential for human health (Hicks et al., 2019; Tacon & Metian,
80 2013). It is predicted that the marine food supply will increase by 21-44 million tonnes by
81 2050, representing a 36-74% increase compared to current yields (Costello et al., 2020)
82 and accounting for one-eighth to one-quarter of the anticipated meat consumption increase,
83 primarily through marine aquaculture expansion (Gentry et al., 2017).

84 Efforts are underway to shift towards a more sustainable and nutritious food system
85 by increasing seafood consumption, as it provides essential micronutrients and contributes
86 to animal protein consumption for 2.9 billion people (W. Willett et al., 2019). Nonetheless,
87 the potential benefits of such a shift could be overshadowed by the risk of seafood
88 contamination resulting from marine pollution incidents, such as the Fukushima Daiichi
89 Nuclear Disaster, the Gulf of Mexico Oil Spill, and the Dalian PetroChina marine pollution
90 incident (K. O. Buesseler, 2012; Fisher et al., 2013; Onda et al., 2020). The impact of these
91 pollution events on marine food availability can range from direct reduction in quantity to

92 the potential for consumer panic due to microbial, chemical, and radioactive contamination
93 (Aruga, 2017a, 2017b; Aruga & Wakamatsu, 2018; Kher et al., 2013). The decline in
94 consumer confidence in food safety in recent years has been linked to food safety incidents
95 (Jonge, Trijp, Renes, & Frewer, 2007; Lobb & Mazzocchi, 2007). Furthermore, consumers'
96 perception of food safety risks will alter their purchase behaviour and ultimate food
97 consumption structure (Grunert, 2005; Yeung & Morris, 2001). While some studies have
98 examined the effects of pollution on food systems, less attention has been given to the
99 changes in food consumption patterns and their nutritional and environmental impacts
100 resulting from changes in consumer behaviour (Feng et al., 2022; Tai, Martin, & Heald,
101 2014; Yates et al., 2021). In developing risk mitigation strategies for food safety, the
102 interests of stakeholders, including consumers, must be prioritized (Kher et al., 2013).

103 On April 13, 2021, the Japanese government announced its decision to release
104 millions of tonnes of filtered and diluted nuclear effluent from the Fukushima Daiichi
105 nuclear power plant into the Pacific Ocean starting after 2023, with an expected 40-year
106 timeline for the full discharge of wastewater (Normile, 2021a, 2021b). The news of Japan's
107 plan to discharge nuclear wastewater into the Pacific Ocean immediately sparked intense
108 discussion in China, mainly about the potential marine pollution and food safety caused by
109 FNWD. China is the world's largest producer, exporter, and processor of aquatic products
110 (FAO, 2022), accounting for 15% of the world's total seafood consumption (FAO, 2021).
111 While some scientists have suggested that the risks are minimal (Nogrady, 2021), consumer
112 perceptions of marine pollution risk could still alter the final food consumption structure.
113 People may choose to abandon seafood or shift to protein-rich aquaculture products,
114 terrestrial-based or aquatic foods from unaffected areas to make up for the loss of protein.

115 In most cases, the intensity of GHG emissions varies among different food products,
116 and livestock products have different GHG emission intensity per unit of protein than
117 marine foods (Clune, Crossin, & Verghese, 2017; Gephart et al., 2021; Nijdam, Rood, &
118 Westhoek, 2012; Parker et al., 2018). Such a food system transition could lead to a change
119 in the nutrition supply and GHG emissions. Given previous experiences with consumer
120 perceptions of radiation-related food safety risks, it is likely that seafood consumption
121 would be impacted if Japan proceeds with its plan to release nuclear wastewater. However,
122 it remains unclear how and to what extent Chinese consumer perceptions of the Fukushima
123 Daiichi nuclear wastewater discharge could affect the food system's nutrition supply and
124 GHG emissions.

125 This study developed an MP-FSRAR model to assess a marine pollution event's
126 potential impacts on food system nutrient supply and GHG emissions. Specifically, we
127 applied the model to investigate how Chinese consumers' risk perceptions of the planned
128 FNWD in 2023 could affect the food system's nutrition supply and GHG emissions. The
129 Materials and Methods section describes the detailed analytical framework and data
130 sources. In brief, the model consists of three components. First, we quantified the supply
131 gaps of eight essential nutrients (protein, calcium, iron, zinc, vitamin A, vitamin B-12,
132 vitamin D, and polyunsaturated fatty acids (PUFA)) under three scenarios of consumer
133 perceptions of seafood safety risks. Next, we estimated the potential supply of protein-rich
134 foods to close protein gaps under four substitution food pathways. We also built a global
135 Protein-based GHG Emission Intensity Database for Terrestrial and Marine Food (PGHG-
136 FOOD) to account for changes in GHG emissions resulting from changes in consumer
137 consumption patterns. Finally, we simulate the potential GHG emissions that could be

138 transferred through the trading chain to other countries under four different substitution
139 food pathways. The model was applied to the planned FNWD, focusing on Chinese
140 consumers as the research subject. We hope this analysis can stimulate a transdisciplinary
141 debate about the impacts of consumer perceptions of marine pollution risks on food system
142 nutrition supply and environmental consequences.

143 **2 Materials and Methods**

144 **2.1 Summary.**

145 The research developed an MP-FSRAR model to analyze the potential impacts of
146 consumer risk perceptions on food system nutrient supply and GHG emissions. Three
147 modules comprised the MP-FSRAR model: (1) Assessing nutrient supply gaps resulting
148 from changes in consumption (without considering alternative food); (2) Estimating the
149 supply of alternative high-protein foods needed to supplement the gaps under four food
150 substitution pathways; (3) Evaluating changes in GHG emissions and the spillover effects
151 on other countries through the food trade. To achieve this, we developed a global Protein-
152 based GHG Emission Intensity Database for Terrestrial and Marine Food (PGHG-FOOD),
153 which includes data for 217 countries (see Supplementary Table).

154 We investigated the potential impact of Japanese nuclear wastewater discharge into
155 the Pacific Ocean on food system nutrient supply and GHG emissions, explicitly
156 concerning Chinese consumers' risk perceptions. We used 2017 as the reference year
157 because it was the most recent year for which data were available from all sources. Using
158 a questionnaire, we examined extreme, middle, and low risk perception scenarios among
159 Chinese consumers toward seafood from different regions worldwide. Based on the
160 questionnaire results, we estimated the nutrient gaps for Chinese consumers and then

161 accounted for alternative food supply and GHG emissions changes resulting from the food
162 system transition to fill the nutrition gap. We also simulated the potential GHG emissions
163 transferred through the trading chain to other countries under four different substitution
164 food pathways. Details of each module and its application are described in the following
165 sections.

166 **2.2 Nutrient supplies gap resulting from changes in consumption.**

167 Consumers' perception of seafood safety risks about exposure and bioaccumulation
168 of radionuclides in marine organisms would change their willingness to consume seafood.
169 This change in willingness would transmit to seafood production, trade, and consumption.
170 We did not include the influence of price because previous research has shown that price
171 is not one of the primary factors consumers consider when confronted with potential food
172 risks. (Aruga, 2017b; Aruga & Wakamatsu, 2018). Consumers' perceived risk of marine
173 pollution on food may be significant, and the willingness to consume would become the
174 critical factor affecting consumption. To characterize consumers' unwillingness to purchase
175 seafood from different regions, we introduced a decline parameter α .

176 To quantify the impact of changes in consumer willingness to consume seafood on
177 regional aquatic production, trade, and consumption, it is necessary to first understand the
178 material flow of seafood in the region and establish the protein supply baseline prior to any
179 changes in consumer behaviour. We utilized the FAO Global Fishery and Aquaculture
180 Production Statistics from FishstatJ and bilateral trade data from UN Comtrade to construct
181 a Marine-based Food Supply Chain Network. We then calculated the nutrition gaps
182 resulting from consumers' decreased willingness to consume seafood by linking the decline
183 parameter, protein content, and seafood consumption. In this study, all aquatic animals,

184 including finfish, crustaceans, and molluscs, harvested from marine (brackish) and
 185 freshwater environments through capture fisheries and aquaculture routes will be
 186 consumed by us, and they are commonly referred to as aquatic products. We excluded
 187 plants, shells, and corals, as well as non-nutritive commodities like ornamental fish, and
 188 estimated the proportion of each commodity that is edible by humans.

189 Aquatic animal food is an inexpensive and essential source of high-quality animal
 190 protein and essential amino acids, omega-3 fatty acids, vitamins, minerals, and trace
 191 elements. These nutrients are crucial for maintaining optimal health and preventing various
 192 diseases. This analysis focused on the eight prominent and specific nutrients in seafood,
 193 including protein, calcium, iron, zinc, vitamin A, vitamin B-12, vitamin D, and PUFA
 194 (Golden et al., 2021; Hicks et al., 2019). To estimate the nutritional value of each seafood
 195 item, we sourced data from the USDA Food Composition Database
 196 (<https://www.ars.usda.gov/northeast-area/beltsville-md-bhnrc/beltsville-human-nutrition-research-center/food-surveys-research-group/docs/fndds-download-databases>).
 197

198 The nutrition gaps in a given country resulting from decreased edible seafood
 199 consumption were quantified using equation (1):

$$200 \quad P_x^n = \sum_{i=1}^i \alpha^{n,w} * Q_i^n * p_{x,i} + \sum_{j=1}^j \sum_{m=1}^m \alpha^{n,w} * T_j^{n,m} * p_{x,j} \quad (1)$$

201 where P_x^n is the nutrition gap in nutrient x for country n (kg). $\alpha^{n,w}$ is the declining
 202 percentage of consumers' willingness in country n to purchase seafood sourced from
 203 country w . Q_i^n is fishing production of food i for domestic consumption of country n (kg).
 204 $p_{x,i}$ is the nutrient x content of the protein-rich food i (kg/kg). $T_j^{n,m}$ is seafood trade
 205 commodity j for domestic consumption from trading partner m (kg). $p_{x,j}$ is the nutrient x
 206 content of the protein-rich commodity j (kg/kg).

207 **2.3 Alternative food supply.**

208 After FNWD, consumers could have the option to supplement their nutrient gaps with
209 aquatic or livestock and crop products. It is important to note that our research operates
210 under the implicit assumption that the total protein requirement for consumers remains
211 unchanged. Researchers often use linear programming models with optimized omnivore,
212 vegan, or novel food diets to estimate the nutritional and environmental changes resulting
213 from the substituted diet (Mazac et al., 2022). In this part, we set up four alternative food
214 supply pathways: 1) the aquatic-based substitution pathway, 2) the current food supply
215 structure substitution pathway, 3) the livestock-based substitution pathway, and 4) the
216 plant-based substitution pathway.

217 In the aquatic-based substitution pathway, consumers who choose to stop eating
218 seafood could be allowed to increase their intake of freshwater aquaculture to supplement
219 their protein gap. Consumers who reduce and maintain their seafood consumption could
220 supplement the protein gap by consuming freshwater aquaculture and imported seafood
221 from the unaffected areas. The proportion of restricted imported seafood in each region
222 depends on how much resistance consumers show towards seafood in that region. This food
223 substitution pathway involves replacing reduced seafood with aquatic products from other
224 sources. The proportion of each source should exclude seafood from areas where
225 consumers are boycotting before replenishing seafood on a business-as-usual basis. In the
226 current food supply structure substitution pathway, the protein gap is filled by eight types
227 of protein-rich foods from terrestrial systems according to each country's current food
228 supply structure, which is carried out on a business-as-usual basis. The aquatic food
229 supplement part is handled in the same way as the aquatic-based substitution pathway. We

230 also establish a livestock-based substitution pathway where livestock products such as
231 beef, pork, chicken, and other meat replenish the protein supply gap. Finally, we provided
232 a plant-based substitution pathway by using pulses (beans, soybeans, peas, nuts, and pulses
233 nes) as the protein substitute which these typical plant-based high protein sources are
234 usually treated as meat protein alternatives (Nijdam et al., 2012).

235 It is important to note that the four protein substitution pathways provide a framework
236 for understanding possible food alternatives in extreme situations rather than an exhaustive
237 list of all possible combinations. The actual food substitutions may vary depending on
238 various factors, including the nutrient content of different foods, dietary culture, consumer
239 income and so on. However, the fluctuations in these values may fall within the upper and
240 lower boundaries of the extreme scenarios.

241 Subsequently, we created a network for a terrestrial-based food supply chain to
242 monitor the regional origins of the substitute food supply. To accomplish this, we utilized
243 the livestock and crops balance tables from FAO
244 (<https://www.fao.org/faostat/en/#data/FBS>) and bilateral trade data from UN Comtrade
245 (<https://comtrade.un.org/data>).

246 **2.4 Impact of GHG emissions and their spillover effects on other countries through** 247 **the food trade.**

248 *Establishing the PGHG-FOOD database.* Under a uniform cradle-to-farm gate
249 accounting boundary, we synthesized multiple data sources to quantify GHG emissions for
250 crops, livestock products, inland aquaculture foods, and marine aquaculture/capture foods.
251 Tab. S1 shows the system boundary definition for marine and terrestrial food GHG
252 emissions. In this study, all food accounting boundaries are from cradle to farm gates.

253 However, seafood, crops, and livestock are subdivided into different accounting methods
254 due to different farming or cultivation approaches. For instance, the farm gate for livestock
255 is grown on the farm until it leaves the farm gate, while the farm gate for marine caught
256 food is defined as when the product is caught ashore.

257 To calculate GHG emissions from crops and livestock products, we used the
258 Comprehensive Accounting of Land-Use Emissions (CALUE) database (Hong et al.,
259 2021), which derives agricultural and land-use change emissions from FAOSTAT
260 (<https://www.fao.org/faostat/en/#data>) and the Bookkeeping of Land Use Emissions model
261 (BLUE) (Hansis, Davis, & Pongratz, 2015). Notably, the CALUE database does not specify
262 GHG emissions related to feed for livestock. We accounted for feed emissions in the GHG
263 emissions of supplemental meat foods, assigning emissions from feed crops to livestock
264 that consume them. We used the Food and Agriculture Biomass Input-Output model
265 (FABIO), a multi-regional supply, use, and input–output table that documents the flows of
266 agricultural and food products in the global economy (Bruckner et al., 2019), to determine
267 the amount of each crop consumed by each livestock. We extracted the domestic and
268 imported quantities of crop consumed by live animals in each country and converted the
269 number of live animal heads into net livestock weight to build a connection between net
270 livestock weight per unit and its crop consumption. We then multiplied the amount of each
271 crop by its GHG emissions intensity to obtain the feed GHG emissions of each country.
272 The FABIO model was up to date through 2014, and we used 2017 food production data
273 to scale their original data equally. The equation (2) can be used to quantify GHG emissions
274 from feed consumed by livestock.

$$275 \quad G_{feed}^n = \sum_{l=1}^l \sum_{f=1}^f (\beta^n * Q_{feed_{l,f}}^n * E_f^n) \quad (2)$$

276 Where G_{feed}^n is the GHG emissions of feed consumed by livestock for country n (kg).
 277 β^n is the ratio of food production in 2017 to 2013 of country n (kg). $Q_{feed_{l,f}}^n$ is the feed f
 278 consumption of livestock l in country n . E_f^n is GHG emissions intensity of feed f for country
 279 n (kg/kg).

280 Only GHG emissions related to freshwater aquaculture were considered for inland
 281 aquatic foods. The marine food category includes marine and brackish water capture or
 282 aquaculture products. The GHG emissions for global freshwater and marine aquaculture
 283 and marine capture fisheries were obtained from MacLeod et al. and Parker et al. research.
 284 GHG emissions intensities for some specific countries were estimated using continent-
 285 average or world-average parameters where data was missing. Finally, a protein-based
 286 GHG intensity inventory for 28 easily accessible protein-rich foods in 217 countries and
 287 regions was developed using food protein content data and mass-weighted GHG intensity.
 288 Please refer to the Supplementary Table for further information.

289 The net food system's GHG emission changes and the food system's GHG emission
 290 absolute changes by the country could be obtained by equation (3):

$$291 \quad G_{net}^n = G_{abs}^n - G_d^n + G_{feed}^n = \sum_{i=1}^i (A_i^n * E_i^n) - \sum_{i=1}^i (P_x^n * E_i^n) + G_{feed}^n \quad (3)$$

292 where G_{net}^n is the net GHG emission changes of the food system in country n (kg).
 293 G_{abs}^n is the food system's GHG emission absolute changes in country n (kg). G_d^n is the
 294 reduced GHG emission due to a decrease in the willingness to consume seafood in country
 295 n (kg). A_i^n is the additional supply of protein of food i in country n (kg). E_i^n is GHG
 296 emission of per-gram protein of food i of country n (g/g). P_x^n is the nutrient gap of food i
 297 (kg). Here, i refers to seafood, and x refers to protein.

298 Processed products must be converted into primary livestock or crops to match the

299 PGHG-FOOD database. To achieve this, we used the technical conversion factor provided
300 by the FAO to convert the first stage of processed products into primary crops and livestock
301 products (Sukhatme, 1960).

302 **2.5 Model application.**

303 Previous studies have shown that a significant proportion of consumers in Japan and
304 its surrounding regions (e.g., Korea), estimated to be around 50-80%, are reported to be
305 unwilling to consume seafood from Fukushima or even from Japan altogether following
306 the Fukushima Daiichi Nuclear Disaster, citing concerns about radiation-related risks
307 (Aruga, 2017a; Kim et al., 2015), due to the potential radiation-related risk. Even when the
308 radioactivity levels of the food are below the official criteria, a considerable proportion of
309 consumers (i.e., 20–50%) still refuse food from Fukushima (Miyata & Wakamatsu, 2018;
310 Ujiie, 2012). To quantify the degree of Chinese consumers' unwillingness to purchase
311 seafood from different regions, we introduced a parameter denoted as $\alpha^{china,w}$. As the
312 FNWD event has not yet occurred, we assessed Chinese consumers' response to seafood
313 through surveys to provide a basis for determining the parameter $\alpha^{china,w}$.

314 *2.5.1 Sampling and sample size.*

315 The data for the sample used to analyze consumers' willingness to consume seafood
316 in China were collected through an online questionnaire survey conducted between March
317 29 and April 5 2022, with participants from 30 provinces of China. A regional stratified
318 sampling method was used to ensure comprehensive country coverage, with random
319 sampling below the provincial level. This sampling method provided good geographic
320 representation and ensured a reasonable proportion of coastal and inland consumers. After
321 a quality check, 2260 valid questionnaires were obtained for the seafood consumption

322 intention survey. The demographic variables related to consumers' age, gender, education,
323 location, family members, number of children, number of elderly people, and income are
324 described in Tab. S2.

325 *2.5.2 Seafood consumption intention survey questionnaire.*

326 *(a) Presurvey.* In the presurvey, Chinese consumers' reactions to Japan's plan to
327 discharge nuclear wastewater into the Pacific Ocean were monitored by Zhejiang public
328 opinion monitoring platform. We collected about 10,000 comments from microblogs,
329 official media, newspapers, forums, news, WeChat, and other platforms on April 12 and its
330 week, and the vast majority of Chinese consumers expressed their unwillingness to
331 consume seafood from the affected waters.

332 *(b) Pre-questionnaire.* The pre-questionnaire was designed based on the presurvey,
333 consumer concerns, and demonstrated tendencies. Some questions and terms in the
334 questionnaire were modified according to respondents' feedback and expert opinions, and
335 the final questionnaire was formalized.

336 *(c) Final questionnaire.* In the final questionnaire, respondents were asked to answer
337 the basic information, daily seafood consumption behaviour, and three parallel sub-
338 questionnaires (A, B, and C) representing low, middle, and extreme consumer reaction
339 scenarios, respectively. In Questionnaire A, the consumers were informed that the
340 discharge of nuclear effluent from the Fukushima Daiichi nuclear power plant into the sea
341 would begin after 2023. In Questionnaire B, respondents were told that radionuclides
342 contained in the nuclear wastewater would spread to the Pacific Ocean. In Questionnaire
343 C, respondents were asked about their responses if radioactive material had been detected
344 in the sea area of region *A* following the discharge of nuclear wastewater from Japan.

345 2.5.3 Data analysis.

346 The data collected from the survey were analyzed to assess consumers' willingness to
347 consume seafood under different scenarios of nuclear wastewater discharge. In
348 Questionnaire A, 34% of respondents said they would no longer consume seafood after the
349 discharge of Japanese nuclear wastewater, 52% said they would reduce their consumption,
350 and 14% said they would keep their consumption the same as usual. In questionnaires B
351 and C, representing middle and extreme consumer reaction scenarios, the proportions of
352 respondents who chose these three options were 46%/63%, 44%/30.0%, and 9%/6%,
353 respectively.

354 To estimate the impact of these consumer responses on seafood consumption, we
355 assumed that consumers who refuse to consume seafood would no longer consume seafood
356 from any region, while those who chose to reduce their seafood consumption would cut
357 half of their usual seafood consumption volume. Very few consumers chose to increase
358 their consumption, so we ignored fluctuations that might arise due to a slight increase in
359 serving size.

360 As a result, $\alpha^{China,w}$ can be calculated by equation (4).

361
$$\alpha^{China,w} = \alpha_1^{China,w} + \alpha_2^{China,w} * 50\% \quad (4)$$

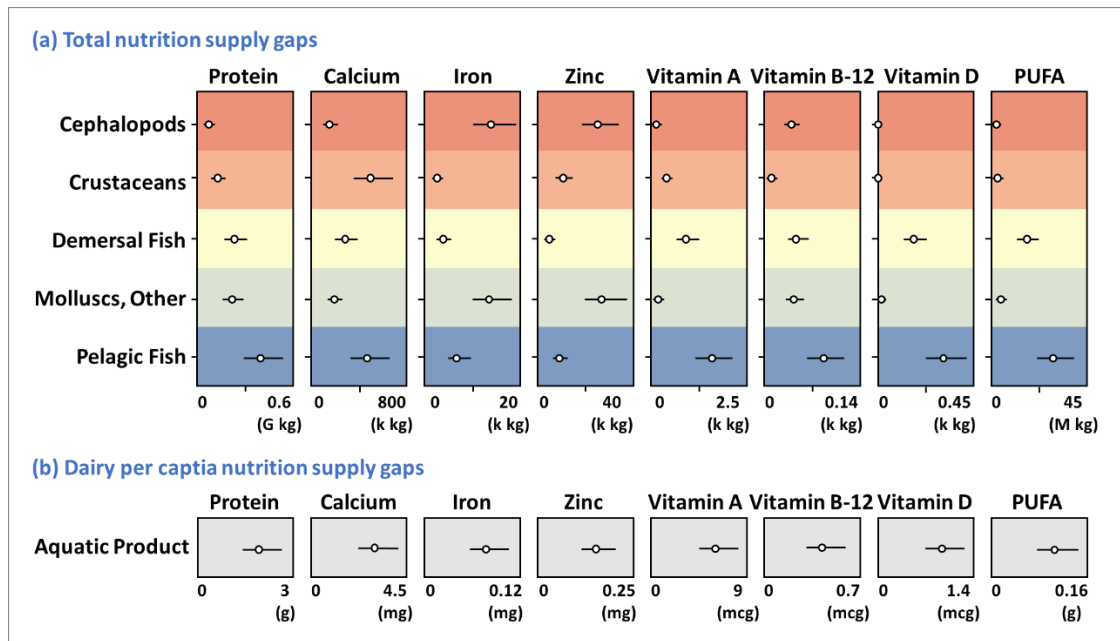
362 $\alpha^{China,w}$ is the declining percentage of Chinses consumers' unwillingness to purchase
363 seafood sourced from region w . $\alpha_1^{China,w}$ was the proportion of Chinses consumers who
364 would refuse to continue consuming seafood from region w . $\alpha_2^{China,w}$ was the proportion
365 of Chinses consumers who would reduce the amount of seafood consumed from region w .
366 The resulting parameter $\alpha^{China,w}$ setting under different Chinese consumers' perceptions
367 of FNWD-induced seafood safety risk is shown in Tab. S3.

368 **3 Results**

369 **3.1 Nutrient supply gaps caused by consumer perceptions of seafood safety risks.**

370 Fig. 1(a) depicts the FNWD-induced total supply gaps for eight typical nutrients
371 (without considering alternative food). We created three scenarios based on Chinese
372 consumers' changed consumption intentions in response to seafood's safety risks. The
373 nutrient gaps from reduced seafood consumption without alternative food supplements are
374 estimated. The results showed that under these scenarios, there would be a potential protein
375 gap of 0.7-1.4 billion kg in China, equivalent to the total protein supply of protein-rich
376 foods in the UK in 2017 (1.4 billion kg). The largest share of the protein deficit was caused
377 by pelagic fish, accounting for 63% of the total protein gap, followed by demersal fish
378 (22%) and molluscs (21%). Cephalopods and molluscs took a higher amount of losses due
379 to their higher zinc and iron content, leading to a mineral gap. A reduction in pelagic fish
380 consumption would result in a 2.9-5.3 k kg vitamin gap, including vitamin A, vitamin B-
381 12, and vitamin D. FNWD was also found to contribute to a gap of 73 million kg PUFA.

382 Fig. 1(b) shows the daily per capita potential nutrient supply gaps under the extreme,
383 middle, and low consumer reaction scenarios. It is estimated that Chinese consumers would
384 suffer a per capita protein gap of 1.5-2.7 g/day. The highest daily per capita nutritional loss
385 among the three minerals is calcium, while vitamin-A showed the largest daily per capita
386 nutritional gap among the vitamins. It should be noted that the nutritional deficit of 80-140
387 mg/day of PUFA resulting from reduced seafood consumption would be difficult to replace
388 with other foods and would lead to an irreparable nutritional loss, as PUFA, mainly n-3
389 PUFA intake, is dominated obtained from marine fish.



390

391 **Fig. 1 Food system potential nutrition gaps in China caused by FNWD.** (a) Total potential nutrition gaps
 392 for five types of seafood caused by FNWD under the low, middle, and extreme consumer reaction scenarios.
 393 The circles represent the middle consumer reaction scenario, while the left and right ends of the horizontal
 394 line represent low and extreme consumer reaction scenarios, respectively. The three scenarios simulate the
 395 consumers' reaction: after the FNWD, the Japanese nuclear wastewater spreads to the Pacific Ocean and the
 396 detection of nuclear radiation in sea A. (b) Daily per capita potential nutrient supply gaps for aquatic
 397 production under the extreme, middle, and low consumer reaction scenarios.

398 **3.2 Growing food demand to close the protein gap.**

399 Fig. 2 illustrates the additional quantity of food that would be required to bridge the
 400 protein gap in the aftermath of the FNWD. We devised four alternative food supply
 401 pathways, including 1) Aquatic-based substitution pathway; 2) Current food supply
 402 structure substitution pathway; 3) Livestock-based substitution pathway; 4) Plant-based
 403 substitution pathway. The details of these food substitution pathways can be found in the
 404 Materials and Methods section. We assumed that after FNWD, consumers might opt for
 405 protein-rich fish from other regions, livestock, or crop products to supplement the protein
 406 gaps. The results showed that the amount of food supplementation required in the extreme

407 and low consumer reaction scenarios would be nearly twofold different. A total of 8.5-15.8
 408 billion kg of aquatic products would be necessary for the aquatic-based substitution
 409 pathways. The demand for freshwater fish constitutes nearly half of the total aquatic food
 410 supply and the supply of molluscs, crustaceans, and demersal fish, each amounting to more
 411 than 1.5 billion kg. In the current food supply structure substitution pathway, the supply of
 412 eight types of protein-rich foods would increase by 7.3-13.5 billion kg. Three main types
 413 of alternatives would be dairy (including milk and eggs) (2.2-4.0 billion kg), pork (1.9-3.5
 414 billion kg), and aquatic products (1.9-3.4 billion kg). In the livestock-based pathway, pork
 415 (4.1-7.6 billion kg) could be the most significant supplementation food, whereas the
 416 demand for chicken (1.4-2.6 billion kg) would also be relatively high. The least amount of
 417 food (3.2-5.9 billion kg) would be required if pulses were used as a protein alternative, with
 418 soybeans accounting for nearly half of the supplement.
 419

		E	M	L			E	M	L
Aquatic-based	Freshwater Fish	7305	5418	3970	Current food supply structure	Pulses	658	488	357
	Demersal Fish	1648	1222	895		Beef	448	332	243
	Pelagic Fish	135	100	73		Other meat	309	229	168
	Marine Fish nes	619	459	336		Pork	3500	2596	1902
	Crustaceans	1846	1369	1003		Chicken	1183	878	643
	Cephalopods	245	182	133		Eggs	1802	1337	979
	Molluscs nes	3776	2800	2052		Milk	2167	1607	1177
	Aquatic Animals nes	229	170	124		Aquatic	3446	2555	1873
	Sum	15803	11720	8586		Sum	13513	10022	7342
		E	M	L			E	M	L
Livestock-based	Beef	971	720	528	Plant-based	Beans	54	40	29
	Other meat	670	497	364		Pulses nes	745	553	405
	Pork	7593	5631	4126		Peas	386	286	210
	Chicken	2567	1904	1395		Nuts	1805	1339	981
	Sum	11801	8752	6413		Soyabeans	2930	2173	1592
					Sum	5920	4391	3217	

420

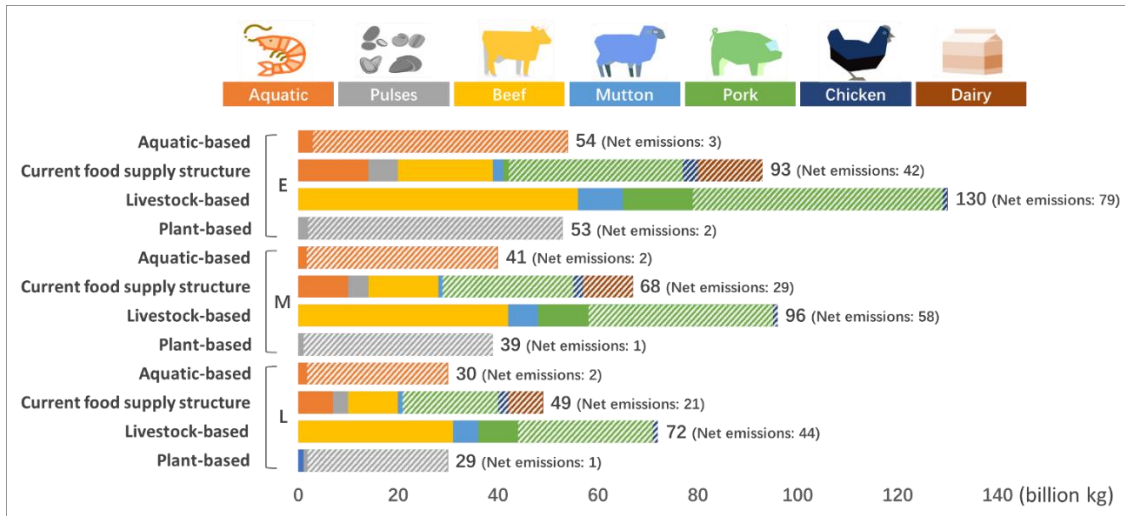
421 **Fig. 2 Additional food supply to close the protein gap under three consumer reaction scenarios and**
 422 **four substitution pathways, million kg.** The transition in a colour gradient from blue to red signifies the
 423 magnitude of food supplementation needed, with darker shades of blue indicating lower quantities and darker

424 shades of red indicating higher quantities. E, M, and L represent extreme, middle, and low consumer reaction
425 scenarios.

426 **3.3 Additional GHG emissions from the food system.**

427 The additional food supplies required to close the protein gap could lead to additional
428 net GHG emissions of 1-79 billion kg CO₂e yr⁻¹ in the food system. Fig. 3 depicts the
429 change in absolute and net GHG emissions in China depending on the consumer reaction
430 scenarios and food substitution pathways selected. The livestock-based substitution
431 pathway would inevitably result in a significant increase in terrestrial GHG emissions,
432 which could reach up to 130 billion kg in the most extreme reaction scenario, equivalent to
433 5.4% of China's terrestrial food system (J. Poore & T. Nemecek, 2018). Pork and beef
434 production are the largest sources of emissions, each contributing over 55 billion kg in the
435 extreme scenario. Compared with the livestock-based substitution pathway (79 billion kg
436 GHG emissions), the current food supply structure substitution pathway could potentially
437 reduce net GHG emissions by almost half (42 billion kg GHG emissions).

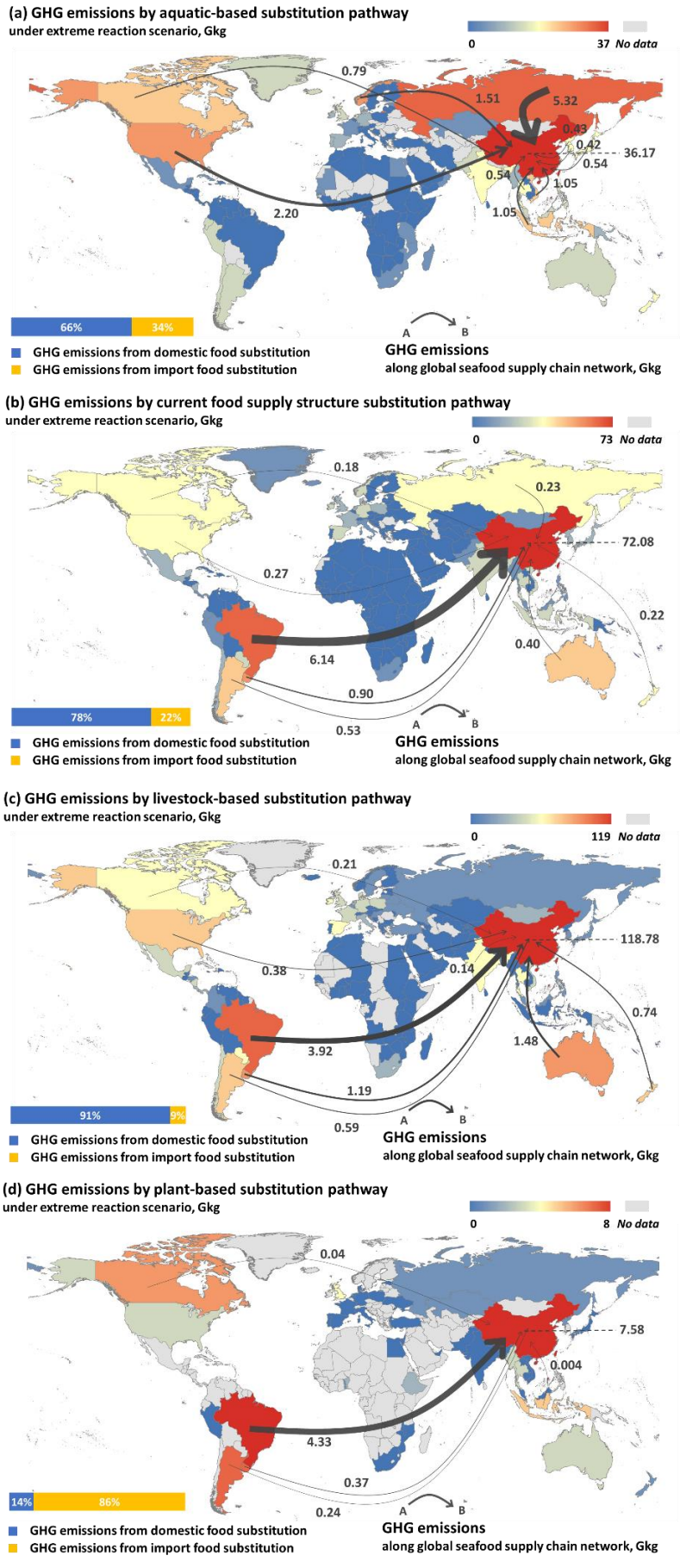
438 In contrast, the aquatic-based and plant-based substitution pathways are more
439 environmentally friendly options. The aquatic-based substitution pathway would lead to
440 fewer potential net GHG emissions of 1-3 billion kg CO₂e yr⁻¹, accounting for only 3.8%
441 of the emissions of the livestock-based substitution pathway. The plant-based substitution
442 pathway would have the least GHG emissions mainly due to the lower GHG emission
443 intensity and higher protein content of pulses. Nevertheless, China's total net GHG
444 emissions under the four food substitution pathways would still be greater than zero,
445 although the impact on GHG emissions resulting after FNWD may be partially mitigated
446 by a shift towards a more plant-based protein diet.



447

448 **Fig. 3** The change in absolute and net GHG emissions in China under different protein supplement pathways,
 449 billion kg. E, M, and L indicate extreme, middle, and low consumer reaction scenarios. The white diagonal
 450 stripe in the figure represents the reduced GHG emissions from declining seafood production.

451 Fig. 4 (a-d) illustrates the changes in GHG emissions of the food system from the
 452 domestic and imported food supply by four substitution pathways under the extreme
 453 reaction scenario. Our results show that more than 60% of the total GHG emissions change
 454 resulting from the aquatic-based, current food supply structure, and livestock-based
 455 substitution pathway would occur domestically, while 86% of the GHG emissions change
 456 from the plant-based substitution pathway would take place abroad. The current food
 457 supply structure, livestock-based, and plant-based food substitution pathways exhibit
 458 similar patterns of GHG emission shifts. Brazil is responsible for the most significant GHG
 459 emissions change, followed by Argentina, Uruguay, and Australia. This is mainly due to
 460 China's large imports of soybeans from Brazil and imports of livestock from Uruguay and
 461 Australia. However, in the aquatic-based food substitution pathway, seafood imports from
 462 Russia would bear the most significant GHG emissions change abroad, generating GHG
 463 emissions of 5.32 billion kg.



465 **Fig. 4** GHG emissions from domestic and imported food substitution by four substitution pathways under the
466 extreme reaction scenario.

467 **4 Discussion**

468 **4.1 Potential nutrition supply challenges and resulting food security threats.**

469 Current assessments of marine pollution risks mainly focus on studies on pollutant
470 diffusion simulation, marine ecosystems, marine life, and human health (Behrens,
471 Schwarzkopf, Lübbecke, & Böning, 2012; K. Buesseler, Aoyama, & Fukasawa, 2011;
472 Fisher et al., 2013; Onda et al., 2020). However, the reassessment of marine pollution
473 incidents on the influence of nutrition supply and GHG emissions along the food supply
474 chain is less considered. Our findings suggest that FNWD would affect domestic food
475 production and international trade as consumer behaviour shifts away from seafood
476 consumption, resulting in multiple nutrient supply gaps. Fish protein is widely known to
477 be rich in lysine, sulfur-containing amino acids, and threonine, making it a healthier source
478 of protein than terrestrial animal proteins (McManus & Newton, 2011). In recent decades,
479 China has made tremendous efforts to increase its protein and micronutrient availability,
480 with per capita seafood consumption increasing sevenfold since 1978 (Bell, Lividini, &
481 Masters, 2021; Fabinyi, 2016). Currently, China's fish protein supply accounts for 22.8%
482 of the total animal protein supply, which exceeds the world average (Barange &
483 *Acuicultura*, 2018).

484 However, in the extreme consumer reaction towards FNWD, Chinese consumers may
485 experience a protein gap of 2.7 g/day due to decreased seafood consumption, equivalent to
486 a 5% loss of daily protein recommendation per person. While some of these nutrients, such
487 as protein, can be partly replenished by terrestrial- and freshwater-based food, others, such
488 as PUFA, are challenging to replenish because the primary dietary source of long-chain

489 omega-3 fatty acids (EPA/DHA) is fisheries (Hamilton, Newton, Auchterlonie, & Müller,
490 2020). The top categories of EPA/DHA -rich animal-source foods are all seafood, including
491 pelagic fish and salmonids (Golden et al., 2021). Our results show that the extreme
492 consumer reaction scenario would result in a 24.7% deficit in daily EPA/DHA
493 recommendation intake for the Chinese consumer. Given the nutritional gaps that may
494 result from changes in consumer behaviour, policies need to ensure adequate nutritional
495 availability to address potential seafood boycotts by Chinese consumers after FNWD.

496 While food substitution can mitigate some of the nutrient gaps, FNWD still reduces
497 food choices for consumers. The diversity of dietary choices is a crucial aspect of food
498 security (Kennedy, Ballard, & Dop, 2011), and decreased seafood selectivity due to FNWD
499 will lead to a rigid and homogeneous food system, which may reduce the safety and
500 flexibility of the food system. Price is an essential factor in food choice (Steenhuis,
501 Waterlander, & de Mul, 2011). However, the price mechanism fails in cases where
502 consumers perceive a food risk, i.e., consumption still decreases when people feel that
503 seafood is contaminated, even if the price is lower. This is similar to the results obtained
504 from previous studies, where consumers are no longer willing to purchase risky foods after
505 perceiving food safety risks, even if the food prices are lower (Aruga, 2017b; Aruga &
506 Wakamatsu, 2018). In addition, this nutritional gap will be more challenging to replenish
507 for consumers in poorer areas, where people have few other options to compensate for
508 these impending micronutrient supply deficiencies (Golden et al., 2016). Communities are
509 often forced to rely on locally harvestable food or less healthy processed foods,
510 compromising micronutrient security in some poor areas, and would pose a whole new
511 challenge to food security in rural China.

512 Compared to aquatic production, terrestrial animal protein production systems are less
513 efficient because they produce food in a two-dimensional space, aggravating land use
514 tension (Napier, Haslam, Olsen, Tocher, & Betancor, 2020). With far less arable land per
515 capita than the world average, China is already facing the enormous challenge of land
516 scarcity (Chen, 2007). In an extreme food substitution pathway, nearly 12 billion kg of
517 livestock products would be needed to replace the protein reduction. Consideration needs
518 to be given to whether there is enough additional arable land, livestock space, and resources
519 in these areas to support additional terrestrial food demand growth. In addition to land use,
520 water, feed, fertilizer, and biodiversity would all be affected by the production of alternative
521 foods (Davis et al., 2016).

522 **4.2 Challenges and potential for GHG mitigation after FNWD.**

523 Different protein supplement pathways after FNWD would result in global GHG
524 emission change patterns. For China, the livestock-based food substitution pathway could
525 increase 44-79 billion kg of potential net GHG emissions, corresponding to 9.7%-17.5%
526 of GHG emissions from the Chinese food system. GHG emissions caused by pork
527 supplementation account for half of the total GHG emissions of all supplemental foods.
528 Red meat, which Chinese consumers prefer, may shift the pressure on terrestrial food
529 systems. China's agricultural production accounts for 13% of global GHG emissions (Zhao
530 et al., 2021). Meeting the consumers' food substitution needs without harming the
531 environment would be a tremendous sustainability challenge for China (Zhao et al., 2021).
532 Our results also found that, in addition to livestock, aquaculture is another crucial source
533 of food nutrient supplementation. Despite differences in the environmental impacts of
534 animal-derived food production sectors, aquaculture almost always produces fewer GHG

535 emissions and uses less land than red meat farming. The large-scale and high-density
536 freshwater and marine aquaculture in China may be a viable way to mitigate the
537 consequences of FNWD through horizontal expansion, intensification, and more efficient
538 resource use (Zhang et al., 2022).

539 Although international food trade could supplement seafood deficits to some extent,
540 it shifts the GHG emissions pressure caused by the agricultural system to other countries
541 and therefore does not reduce net GHG emissions from the global food system. China
542 would import more seafood from Russia, Norway, etc., after FNWD to close the protein
543 gap, which will also bring additional GHG emissions from these countries. Brazil,
544 Argentina, and Uruguay in South America bear most of the GHG emissions increase from
545 livestock and plant-based food substitution in China. Using comparative environmental
546 advantages, potential ways to mitigate FNWD could be found through international trade.
547 For example, China imports fewer emissions-intensive soybeans from the United States
548 than Brazil and Argentina, suggesting that global emissions might be avoided if the United
549 States could produce more soybeans and export them to China (Hong et al., 2022).

550 However, consumers' concerns about seafood safety may further emphasize the
551 necessity of domestically produced land-based food and uncontaminated seafood products
552 after Japan's nuclear wastewater discharge. The increased demand for these products may
553 pressure those producing countries to restrict exports to ensure domestic supply as
554 evidenced by many previous cases (Falkendal et al., 2021; Reinhart, 2020). While export
555 restrictions may be understandable from the perspective of domestic supply, they could
556 have severe implications for global food security. Such food export restrictions could push
557 up food prices and exacerbate hunger and poverty (Carriquiry, Dumortier, & Elobeid, 2022;

558 Chung & Liu, 2022). It could also intensify competition between China and other countries
559 between land-based and uncontaminated seafood products.

560 **4.3 Uncertainty analysis**

561 Our study provides valuable insights into the potential effects of FNWD on both
562 domestic and international food systems, but it is essential to recognize the inherent
563 uncertainties associated with the model used. The focus of this study is to evaluate the
564 consumers' perceived changes after the discharge of Japanese nuclear wastewater and its
565 possible nutrition and climate consequences, but our method may not capture all types of
566 impacts. Our analysis assumes a static relationship between nutrient supply and consumer
567 behaviour, whereas, in reality, this relationship may be dynamic and influenced by various
568 social and economic factors. Given the complexity and variability of consumer behaviour,
569 the boycott's scale, duration, and extent of consumer substitution are challenging to predict
570 accurately. Hence, our estimates based on intentions may overestimate actual consumption
571 levels. However, if real-world data on production or consumption reductions become
572 available, our model can be updated accordingly for more precise results. In addition, the
573 impacts of food substitution on nutrient availability may vary depending on the specific
574 foods substituted and the geographic region in question, and the actual substitution
575 pathways may be more complex and diverse than our model.

576 We must acknowledge that the issue of nuclear wastewater is highly controversial and
577 sensitive, and it may attract widespread attention and public concern. Due to consumers'
578 cognitive biases, when hazards are perceived as unknown, people tend to evaluate hazards
579 as more threatening. People prefer reducing current risks and stricter regulatory measures
580 for uncontrollable, global catastrophic, and difficult-to-reduce risk factors, such as nuclear

581 technology and radioactive waste. Thus, caution is needed when interpreting our study's
582 findings, and further research is necessary to understand better the intricate and dynamic
583 interactions between consumer behaviour, nutrient supply, and GHG emissions.

584 **5 Conclusions**

585 The announcement of the FNWD news has led to widespread discussion among
586 Chinese consumers. To investigate the potential impact of FNWD caused by the change in
587 Chinese consumers' willingness to protein supply and GHG emissions, we developed an
588 MP-FSRAR model. We found out that when a marine environmental risk incident occurs,
589 the changes in food production due to changes in people's consumption preferences will
590 result in a wide-ranging impact on nutrition security and GHG emissions mitigation efforts,
591 which are much broader than previously assumed. Recognizing the possible larger scale
592 and profound indirect impact of FNWD can give timely warnings to reduce risks and allow
593 for better coordinated regional responses to the incident.

594 Considering the multi-risks that FNWD action may bring, we call on all parties
595 concerned to review the FNWD action carefully and jointly formulate acceptable nuclear
596 wastewater treatment schemes. First, countries should guide the public's food consumption
597 choices, promote the diversification of protein food sources, and ensure a stable supply of
598 high-quality protein sources to secure a food supply based on balanced nutrition. Second,
599 keeping abreast of changes in consumers' dietary structure to monitor the pressure on the
600 terrestrial food system and the transition to a diet with less climate impact can increase
601 production efficiency to keep up with the change in human demand. Finally, we should
602 strengthen the ecological environment monitoring of relevant sea areas, carry out
603 simulation studies on the impact of nuclear wastewater pollution, assess potential risks, and

604 formulate relevant countermeasures.

605 **Data availability**

606 The population is available from the World Bank database
607 (<https://databank.worldbank.org/indicator/SP.POP.TOTL/1ff4a498/Popular-Indicators>).

608 FAO data on protein supply and agricultural GHG emissions for crop and livestock
609 production in 2017 (<http://www.fao.org/faostat/en/#data>). CALUE data on GHG emissions

610 for crops and livestock are referred from Hong et al. (Hong et al., 2021). The GHG
611 emissions for global freshwater and marine aquaculture and marine capture fisheries were

612 sourced from MacLeod et al. (MacLeod, Hasan, Robb, & Mamun-Ur-Rashid, 2020) and
613 Parker et al. (Parker et al., 2018). FABIO data on livestock's feed of multi-regional physical

614 supply-use and input-output tables covering global agriculture and forestry
615 (https://github.com/martinbruckner/fabio_v1). The nutrition value of each seafood was

616 obtained from USDA Food Composition Database ([https://www.ars.usda.gov/northeast-](https://www.ars.usda.gov/northeast-area/beltsville-md-bhnrc/beltsville-human-nutrition-research-center/food-surveys-)
617 [area/beltsville-md-bhnrc/beltsville-human-nutrition-research-center/food-surveys-](https://www.ars.usda.gov/northeast-area/beltsville-md-bhnrc/beltsville-human-nutrition-research-center/food-surveys-)

618 [research-group/docs/fndds-download-databases/](https://www.ars.usda.gov/northeast-area/beltsville-md-bhnrc/beltsville-human-nutrition-research-center/food-surveys-research-group/docs/fndds-download-databases/)). FishStatJ data on inland and marine
619 aquaculture and capture foods production in 2017

620 (<https://www.fao.org/fishery/en/statistics/software/fishstatj>). UN Comtrade data on fishery
621 products trade data (<https://comtrade.un.org/data>). The R original code used in data

622 processing is available from the author on request.

623 **Competing Interest Statement**

624 All other authors declare they have no competing interests.

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630 **Author Contributions:**

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634 **and Mingxing Sun:** Methodology, Formal analysis, Writing, Validation. **Jiafu Mao,**
635 **Mingzhou Jin, Cecilia M. Villas Bôas de Almeida, Anthony Chiu, Lan Yang, Linxiu**
636 **Zhang, Chun Ding:** Formal analysis, Writing, Validation. **Yutao Wang:** Conceptualization,
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639 **References**

- 640 Aruga, K. (2017a). *Consumer Reaction, Food Production and the Fukushima Disaster*.
641 Aruga, K. (2017b). Consumer responses to food produced near the Fukushima nuclear plant.
642 *Environmental Economics and Policy Studies*, 19(4), 677-690. doi:10.1007/s10018-016-
643 0169-y
644 Aruga, K., & Wakamatsu, H. (2018). Consumer Perceptions toward Seafood Produced near the
645 Fukushima Nuclear Plant. *Marine Resource Economics*, 33, 000-000. doi:10.1086/698998
646 Barange, M. J. F. y. F., & Acuicultura, A. S. F. A. S. d. P. e. d. I. A. F. A. E. d. P. y. (2018). Fishery
647 and aquaculture statistics. 1-82.
648 Behrens, E., Schwarzkopf, F. U., Lübbecke, J. F., & Böning, C. W. (2012). Model simulations on
649 the long-term dispersal of 137 Cs released into the Pacific Ocean off Fukushima.
650 *Environmental Research Letters*, 7(3), 034004. doi:10.1088/1748-9326/7/3/034004
651 Bell, W., Lividini, K., & Masters, W. A. (2021). Global dietary convergence from 1970 to 2010
652 altered inequality in agriculture, nutrition and health. *Nature Food*, 2(3), 156-165.
653 doi:10.1038/s43016-021-00241-9
654 Bruckner, M., Wood, R., Moran, D., Kuschnig, N., Wieland, H., Maus, V., & Börner, J. (2019).
655 FABIO — The Construction of the Food and Agriculture Biomass Input–Output Model.
656 *Environmental Science & Technology*, 53. doi:10.1021/acs.est.9b03554
657 Buesseler, K., Aoyama, M., & Fukasawa, M. (2011). Impacts of the Fukushima nuclear power
658 plants on marine radioactivity. *Environmental science technology*, 45(23), 9931-9935.
659 Buesseler, K. O. (2012). Fishing for answers off Fukushima. *Science*, 338(6106), 480-482.

660 Carriquiry, M., Dumortier, J., & Elobeid, A. (2022). Trade scenarios compensating for halted
661 wheat and maize exports from Russia and Ukraine increase carbon emissions without
662 easing food insecurity. *Nature Food*, 3(10), 847-850. doi:10.1038/s43016-022-00600-0
663 Chen, J. (2007). Rapid urbanization in China: A real challenge to soil protection and food
664 security. *CATENA*, 69(1), 1-15. doi:<https://doi.org/10.1016/j.catena.2006.04.019>
665 Chung, M. G., & Liu, J. (2022). International food trade benefits biodiversity and food security in
666 low-income countries. *Nature Food*, 3(5), 349-355. doi:10.1038/s43016-022-00499-7
667 Clark, M. A., Domingo, N. G. G., Colgan, K., Thakrar, S. K., Tilman, D., Lynch, J., . . . Hill, J. D.
668 (2020). Global food system emissions could preclude achieving the 1.5° and 2°C climate
669 change targets. *Science*, 370(6517), 705. doi:10.1126/science.aba7357
670 Clune, S., Crossin, E., & Verghese, K. (2017). Systematic review of greenhouse gas emissions for
671 different fresh food categories. *Journal of Cleaner Production*, 140, 766-783.
672 doi:10.1016/j.jclepro.2016.04.082
673 Costello, C., Cao, L., Gelcich, S., Cisneros-Mata, M. A., Free, C. M., Froehlich, H. E., . . .
674 Lubchenco, J. (2020). The future of food from the sea. *Nature*, 588(7836), 95-100.
675 doi:10.1038/s41586-020-2616-y
676 Crippa, M., Solazzo, E., Guizzardi, D., Monforti-Ferrario, F., Tubiello, F. N., & Leip, A. (2021). Food
677 systems are responsible for a third of global anthropogenic GHG emissions. *Nature*
678 *Food*, 2(3), 198-209. doi:10.1038/s43016-021-00225-9
679 Davis, K., Gephart, J., Emery, K., Leach, A., Galloway, J., & D'Odorico, P. (2016). Meeting future
680 food demand with current agricultural resources. *Global Environmental Change*, 39,
681 125-132. doi:10.1016/j.gloenvcha.2016.05.004
682 Fabinyi, M. (2016). Sustainable seafood consumption in China. *Marine Policy*, 74, 85-87.
683 doi:10.1016/j.marpol.2016.09.020
684 Falkendal, T., Otto, C., Schewe, J., Jägermeyr, J., Konar, M., Kummu, M., . . . Puma, M. J. (2021).
685 Grain export restrictions during COVID-19 risk food insecurity in many low- and middle-
686 income countries. *Nature Food*, 2(1), 11-14. doi:10.1038/s43016-020-00211-7
687 FAO. (2021). *FAO Yearbook of Fishery and Aquaculture Statistics*. Rome, Italy: FAO.
688 FAO. (2022). *The State of World Fisheries and Aquaculture 2022. Towards Blue Transformation*.
689 Rome, Italy: FAO.
690 Feng, Z., Xu, Y., Kobayashi, K., Dai, L., Zhang, T., Agathokleous, E., . . . Yue, X. (2022). Ozone
691 pollution threatens the production of major staple crops in East Asia. *Nature Food*, 3(1),
692 47-56. doi:10.1038/s43016-021-00422-6
693 Fisher, N. S., Beaugelin-Seiller, K., Hinton, T. G., Baumann, Z., Madigan, D. J., & Garnier-Laplace,
694 J. (2013). Evaluation of radiation doses and associated risk from the Fukushima nuclear
695 accident to marine biota and human consumers of seafood. *Proceedings of the National*
696 *Academy of Sciences*, 110(26), 10670. doi:10.1073/pnas.1221834110
697 Foley, J. A., Ramankutty, N., Brauman, K. A., Cassidy, E. S., Gerber, J. S., Johnston, M., . . . Zaks, D.
698 P. (2011). Solutions for a cultivated planet. *Nature*, 478(7369), 337-342.
699 doi:10.1038/nature10452
700 Fujimori, S., Hasegawa, T., Krey, V., Riahi, K., Bertram, C., Bodirsky, B. L., . . . van Vuuren, D.
701 (2019). A multi-model assessment of food security implications of climate change
702 mitigation. *Nature Sustainability*, 2(5), 386-396. doi:10.1038/s41893-019-0286-2
703 Gentry, R. R., Froehlich, H. E., Grimm, D., Kareiva, P., Parke, M., Rust, M., . . . Halpern, B. S.
704 (2017). Mapping the global potential for marine aquaculture. *Nat Ecol Evol*, 1(9), 1317-
705 1324. doi:10.1038/s41559-017-0257-9
706 Gephart, J. A., Henriksson, P. J., Parker, R. W., Shepon, A., Gorospe, K. D., Bergman, K., . . .
707 Hornborg, S. (2021). Environmental performance of blue foods. *Nature*, 597, 360-365.

708 Godfray, H. C. J., Beddington, J. R., Crute, I. R., Haddad, L., Lawrence, D., Muir, J. F., . . . Toulmin,
709 C. (2010). Food security: the challenge of feeding 9 billion people. *Science*, 327(5967),
710 812-818.

711 Golden, C. D., Allison, E. H., Cheung, W. W. L., Dey, M. M., Halpern, B. S., McCauley, D. J., . . .
712 Myers, S. S. (2016). Nutrition: Fall in fish catch threatens human health. *Nature*,
713 534(7607), 317-320. doi:10.1038/534317a

714 Golden, C. D., Koehn, J. Z., Shepon, A., Passarelli, S., Free, C. M., Viana, D. F., . . . Thilsted, S. H.
715 (2021). Aquatic foods to nourish nations. *Nature*, 598(7880), 315-320.
716 doi:10.1038/s41586-021-03917-1

717 Grunert, K. G. (2005). Food quality and safety: consumer perception and demand. *European*
718 *Review of Agricultural Economics*, 32(3), 369-391. doi:10.1093/eurrag/jbi011

719 Hamilton, H. A., Newton, R., Auchterlonie, N. A., & Müller, D. B. (2020). Systems approach to
720 quantify the global omega-3 fatty acid cycle. *Nature Food*, 1(1), 59-62.
721 doi:10.1038/s43016-019-0006-0

722 Hansis, E., Davis, S. J., & Pongratz, J. (2015). Relevance of methodological choices for accounting
723 of land use change carbon fluxes. *Global Biogeochemical Cycles*, 29(8), 1230-1246.
724 doi:<https://doi.org/10.1002/2014GB004997>

725 Hicks, C. C., Cohen, P. J., Graham, N. A. J., Nash, K. L., Allison, E. H., D’Lima, C., . . . MacNeil, M. A.
726 (2019). Harnessing global fisheries to tackle micronutrient deficiencies. *Nature*,
727 574(7776), 95-98. doi:10.1038/s41586-019-1592-6

728 Hong, C., Burney, J. A., Pongratz, J., Nabel, J. E. M. S., Mueller, N. D., Jackson, R. B., & Davis, S. J.
729 (2021). Global and regional drivers of land-use emissions in 1961–2017. *Nature*,
730 589(7843), 554-561. doi:10.1038/s41586-020-03138-y

731 Hong, C., Zhao, H., Qin, Y., Burney, J. A., Pongratz, J., Hartung, K., . . . Davis, S. J. (2022). Land-use
732 emissions embodied in international trade. *Science*, 376(6593), 597-603.
733 doi:10.1126/science.abj1572

734 Jonge, J., Trijp, H., Renes, R., & Frewer, L. (2007). Understanding Consumer Confidence in the
735 Safety of Food: Its Two-Dimensional Structure and Determinants. *Risk analysis : an*
736 *official publication of the Society for Risk Analysis*, 27, 729-740. doi:10.1111/j.1539-
737 6924.2007.00917.x

738 Kennedy, G., Ballard, T., & Dop, M. C. (2011). *Guidelines for measuring household and individual*
739 *dietary diversity*: Food and Agriculture Organization of the United Nations.

740 Kher, S., Jonge, J., Wentholt, M., Deliza, R., Andrade, J., Cnossen, H., . . . Frewer, L. (2013).
741 Consumer perceptions of risks of chemical and microbiological contaminants associated
742 with food chains: A cross-national study. *International Journal of Consumer Studies*, 37,
743 73-83.

744 Kim, N. H., Cho, T. J., Kim, Y. B., Park, B. I., Kim, H. S., & Rhee, M. S. (2015). Implications for
745 effective food risk communication following the Fukushima nuclear accident based on a
746 consumer survey. *Food Control*, 50, 304-312. doi:10.1016/j.foodcont.2014.09.008

747 Laborde, D., Martin, W., Swinnen, J., & Vos, R. (2020). COVID-19 risks to global food security.
748 *Science*, 369(6503), 500-502. doi:10.1126/science.abc4765

749 Lobb, A., & Mazzocchi, M. (2007). Domestically produced food: Consumer perceptions of origin,
750 safety and the issue of trust. *Food Economics - Acta Agriculturae Scandinavica, Section C*,
751 4, 3-12. doi:10.1080/16507540701192485

752 MacLeod, M. J., Hasan, M. R., Robb, D. H. F., & Mamun-Ur-Rashid, M. (2020). Quantifying
753 greenhouse gas emissions from global aquaculture. *Scientific Reports*, 10(1), 11679.
754 doi:10.1038/s41598-020-68231-8

755 Mazac, R., Meinilä, J., Korkalo, L., Järviö, N., Jalava, M., & Tuomisto, H. L. (2022). Incorporation
756 of novel foods in European diets can reduce global warming potential, water use and
757 land use by over 80%. *Nature Food*, 3(4), 286-293. doi:10.1038/s43016-022-00489-9
758 McManus, A., & Newton, W. (2011). Seafood, nutrition and human health: A synopsis of the
759 nutritional benefits of consuming seafood.

760 Miyata, T., & Wakamatsu, H. (2018). Who refuses safe but stigmatized marine products due to
761 concern about radioactive contamination? *Fisheries Science*, 84(6), 1119-1133.
762 doi:10.1007/s12562-018-1250-1

763 Mueller, N. D., Gerber, J. S., Johnston, M., Ray, D. K., Ramankutty, N., & Foley, J. A. (2012).
764 Closing yield gaps through nutrient and water management. *Nature*, 490(7419), 254-
765 257. doi:10.1038/nature11420

766 Napier, J., Haslam, R., Olsen, R., Tocher, D., & Betancor, M. (2020). Agriculture can help
767 aquaculture become greener. *Nature Food*, 1, 680-683. doi:10.1038/s43016-020-00182-
768 9

769 Nature, N. E. J. (2022). The war in Ukraine is exposing gaps in the world's food-systems
770 research. *604(7905)*, 217-218.

771 Nijdam, D., Rood, T., & Westhoek, H. (2012). The price of protein: Review of land use and carbon
772 footprints from life cycle assessments of animal food products and their substitutes.
773 *Food Policy*, 37(6), 760-770. doi:10.1016/j.foodpol.2012.08.002

774 Nogrady, B. (2021). Scientists OK plan to release one million tonnes of waste water from
775 Fukushima. *Nature News*. doi:<https://doi.org/10.1038/d41586-021-01225-2>

776 Normile, D. (2021a). Japan plans to release Fukushima's wastewater into the ocean [Press
777 release]

778 Normile, D. (2021b). Japan plans to release Fukushima's wastewater into the ocean. *Science*
779 *News*. doi:10.1126/science.abi9880

780 Onda, Y., Taniguchi, K., Yoshimura, K., Kato, H., Takahashi, J., Wakiyama, Y., . . . Smith, H. (2020).
781 Radionuclides from the Fukushima Daiichi Nuclear Power Plant in terrestrial systems.
782 *Nature Reviews Earth & Environment*, 1(12), 644-660. doi:10.1038/s43017-020-0099-x

783 Organization, W. H. (2020). *The state of food security and nutrition in the world 2020:*
784 *transforming food systems for affordable healthy diets* (Vol. 2020): Food & Agriculture
785 Org.

786 Parker, R. W. R., Blanchard, J. L., Gardner, C., Green, B. S., Hartmann, K., Tyedmers, P. H., &
787 Watson, R. A. (2018). Fuel use and greenhouse gas emissions of world fisheries. *Nature*
788 *Climate Change*, 8(4), 333-337. doi:10.1038/s41558-018-0117-x

789 Poore, J., & Nemecek, T. (2018). Reducing food's environmental impacts through producers and
790 consumers. *Science*, 360(6392), 987-992.

791 Poore, J., & Nemecek, T. (2018). Reducing food's environmental impacts through producers and
792 consumers. *Science*, 360(6392), 987-992. doi:10.1126/science.aaq0216

793 Reinhart, C. M. (2020). How can we prevent a COVID-19 food crisis? Retrieved from
794 <https://www.weforum.org/agenda/2020/05/preventing-a-covid-19-food-crisis/>

795 Smith Martin, D., Roheim Cathy, A., Crowder Larry, B., Halpern Benjamin, S., Turnipseed, M.,
796 Anderson James, L., . . . Selkoe Kimberly, A. (2010). Sustainability and Global Seafood.
797 *Science*, 327(5967), 784-786. doi:10.1126/science.1185345

798 Steenhuis, I. H. M., Waterlander, W. E., & de Mul, A. (2011). Consumer food choices: the role of
799 price and pricing strategies. *Public Health Nutrition*, 14(12), 2220-2226.
800 doi:10.1017/S1368980011001637

801 Sukhatme, P. (1960). *Technical conversion factors for agricultural commodities*: FAO.

802 Tacon, A., & Metian, M. (2013). Fish Matters: Importance of Aquatic Foods in Human Nutrition
803 and Global Food Supply. *Reviews in Fisheries Science*, 21.
804 doi:10.1080/10641262.2012.753405

805 Tai, A. P. K., Martin, M. V., & Heald, C. L. (2014). Threat to future global food security from
806 climate change and ozone air pollution. *Nature Climate Change*, 4(9), 817-821.
807 doi:10.1038/nclimate2317

808 Tilman, D., Balzer, C., Hill, J., & Befort, B. L. (2011). Global food demand and the sustainable
809 intensification of agriculture. *Proceedings of the National Academy of Sciences* 108(50),
810 20260-20264. doi:10.1073/pnas.1116437108

811 Ujiie, K. (2012). Consumer's Evaluation on Radioactive Contamination of Agricultural Products in
812 Japan: Decomposition of WTA into a Part Due to Radioactive Contamination and a Part
813 Due to Area of Origin. *Journal of Food System Research*, 19(2), 142-155.
814 doi:10.5874/jfsr.19.142

815 Willett, W., Rockström, J., Loken, B., Springmann, M., Lang, T., Vermeulen, S., . . . Murray, C. J. L.
816 (2019). Food in the Anthropocene: the EAT-Lancet Commission on healthy diets from
817 sustainable food systems. (1474-547X (Electronic)).

818 Willett, W., Rockström, J., Loken, B., Springmann, M., Lang, T., Vermeulen, S., . . . Wood, A. J. T. I.
819 (2019). Food in the Anthropocene: the EAT–Lancet Commission on healthy diets from
820 sustainable food systems. 393(10170), 447-492.

821 Yates, J., Deeney, M., Rolker, H. B., White, H., Kalamatianou, S., & Kadiyala, S. (2021). A
822 systematic scoping review of environmental, food security and health impacts of food
823 system plastics. *Nature Food*, 2(2), 80-87. doi:10.1038/s43016-021-00221-z

824 Yeung, R. M. W., & Morris, J. (2001). Food safety risk: Consumer perception and purchase
825 behaviour. *British Food Journal*, 103(3), 170-187. doi:10.1108/00070700110386728

826 Zhang, W., Belton, B., Edwards, P., Henriksson, P. J. G., Little, D. C., Newton, R., & Troell, M.
827 (2022). Aquaculture will continue to depend more on land than sea. *Nature*, 603(7900),
828 E2-E4. doi:10.1038/s41586-021-04331-3

829 Zhao, H., Chang, J., Havlík, P., van Dijk, M., Valin, H., Janssens, C., . . . Obersteiner, M. (2021).
830 China's future food demand and its implications for trade and environment. *Nature*
831 *Sustainability*, 4(12), 1042-1051. doi:10.1038/s41893-021-00784-6

832